

Assessment of Pesticide Residues in Fish Tissues and Water at Various Time of the Year

Karzan Namiq^{1,2*}, Nuha Hameed Albassam¹, Nihad Khurshid Wahab¹

¹Department of Animal Production, College of Agriculture, University of Tikrit, Salah Al-Deen, Iraq.

²Sulaimani Polytechnic University, Sulaymaniyah 46001, Iraq.

*Corresponding author: karzan.namiq@spu.edu.iq

Article history:

Received: 17 February 2025

Accepted: 19 May 2025

Published: 30 June 2025

Keywords: Fish tissues, Pesticide, Dukan Lake, Heat treatment, Water pollution.

Abstract

The present study aimed to assess three pesticide residues (hexachlorobenzene, cypermethrin and chlorpyrifos) in the muscle of common carp (*Cyprinus carpio*), barbus grypus (*Arabibarbus grypus*) and water from Dukan Lake. About 60 samples were collected from different sites on the Dukan Lake in Iraq from different seasons (November 2023 to August 2024). Fish samples were analyzed fresh and grilled for the presence of pesticide residues using the QuEChERS method for extraction and water extraction was done by solid-phase extraction method accompanied by cleanup and detection using GC–MS (gas chromatography-mass spectrometry). The results revealed that the examined pesticides were identified in both types of fish muscles and water during different seasons, except hexachlorobenzene, which was not discovered in either fish type in the spring and in water during the summer. All pesticide residues were less than the maximum residue limit (MRL) in Japan, and just one sample exceeded the maximum residue limit in the European Union. Cooking (grilling) fish samples lowered pesticide residual levels to at least 25% of those measured in raw muscle. The trial findings indicated that eating common carp and barbus grypus and drinking water provides minor health hazards owing to residues of the three studied pesticides.

<https://dx.doi.org/10.52951/dasj.25170114>

This article is open-access under the CC BY 4.0 license (<http://creativecommons.org/licenses/by/4.0/>).

Introduction

One of the many inconsistencies in modern civilization is the use of pesticides to repel pests from plants. Pesticides are a diverse set of substances with established toxicity and catastrophic consequences for humans, animals, and the environment; nonetheless, they are widely employed worldwide in agriculture and health programs (Kumari *et al.*, 2014). Pesticides harm a diverse range of non-target creatures, including aquatic invertebrates and fish (Burkepile *et al.*, 2000). Pesticides are compounds that can pollute both aquatic and terrestrial habitats. Pesticides sprayed in one location can end up in the aquatic environment via drift, leaching, and drainage (Srivastava *et al.*, 2010).

Fish can consume tainted food or directly absorb OCPs from the water. The area where pesticides build up in fish varies depending on the absorption pathway (Monirith *et al.*, 2000). Furthermore, the quantity of OCPs in water is reflected in the amount found in the gills, whereas the liver stores all pesticides. Thus, fish may be one of the important bioindicators of pesticide contamination in freshwater environments (Rajendran *et al.*, 2005; Ntow, 2005).

Since fish accumulate pollutants from aquatic environment, the habitat of aquatic species, eating them might be seen as one of the main causes of worry. Over time, pesticides bioaccumulate in aquatic creatures' tissue (Jiries *et al.*, 2002; Yahia and Elsharkawy, 2014; Hassan *et al.*, 2024).

Eating fish may be one of the main ways that people are exposed to environmental pollutants (Storelli, 2008; Yahia and Elsharkawy, 2014; Hassan *et al.*, 2025). Aquaculturists may suffer financially due to the buildup of pesticides, which may lower fish quality in hatcheries and jeopardize their survival after release (Storelli, 2008; Botaro *et al.*, 2011).

Understanding the dispersion and permanence of pesticides in the environment is necessary for their efficient and successful usage. In order to determine dispersion patterns during application and to assess their environmental fate, a significant portion of this knowledge is gathered through the collection and analysis of soil and water samples (Zidan *et al.*, 1997). According to VanCuren (2003) and Maule *et al.* (2007) pesticides can attach themselves to the soil and then reach the aquatic environment by precipitation runoff or dust carried by the air. The current study's objectives are to evaluate the levels of three pesticide residues in fish tissues (common carp and barbus grypus) and water from Dukan Lake at various times of the year and investigate how heat treatment affects these levels.

Materials and methods

Ethics Approval

The scientific ethical committee of Animal Production Department, College of Agriculture, University of Tikrit, approved this study and give the Ethical Number 3/7/1517 (27/11/2023), all applicable national and international guidelines for the care and use of animals were followed.

Sample collection

Water and muscle samples were collected from different sites on Dukan Lake in Iraq from different seasons (November 2023 to August 2024). The study and sample collection were conducted in Sulaimani, Iraq. The absence of the investigated pesticides was verified by testing the blank samples. Analyzed blank samples that were discovered to be devoid of target analytes were utilized as blanks and to create standard curves that matched the matrix. The remaining fraction was spiked and utilized for recovery, standard calibration, and sensitivity investigations after an aliquot of the samples was used for the selectivity research.

Chemical and apparatus

The standards for pesticides were; hexachlorobenzene (HCB) (98%), cypermethrin (CMT) (94%), and chlorpyrifos (CPF) (96%), Merck Ltd. provided acetonitrile (ACN) HPLC grade, acetic acid (99.9%), primary secondary amine (PSA) with a particle size of 40 μm , octadecylsilane (C18, 50 μm), sodium chloride (NaCl), anhydrous magnesium sulphate (MgSO_4), and syringe filters (0.45 μm). The injector, interface, and ion source were heated to 250 $^{\circ}\text{C}$, and split less injection (1.0 min) was carried out using helium as the carrier gas at a rate of flow of 5 mL/min. The oven temperature was set at 120 and raised at a rate of 20 $^{\circ}\text{C}$ per minute.

The injector's temperature was adjusted to 250 $^{\circ}\text{C}$, and a splitless injection (1.0 min) was done using helium as the carrier gas at a flow rate of 5 mL/min. The oven temperature was set to 130 $^{\circ}\text{C}$ and raised at a rate of 10 $^{\circ}\text{C}/\text{min}$ for 2 minutes from 130 to 200 $^{\circ}\text{C}$. The temperature was then raised from 32 $^{\circ}\text{C}/\text{min}$ to 270 $^{\circ}\text{C}$ for 3 minutes and kept for 4 minutes. The injection volume was 50 μL using a split-less injection mode (Sartarelli *et al.*, 2012).

Gas chromatography-mass spectrometry (GC-MS)

Shimadzu gas chromatograph 17 A with an ECD (electron capture detection), the primary compound was separated. Column: Supelco Analytical Co., UK, supplied capillary columns, 30 m DB-5, having an internal diameter of 0.25 mm and a thickness of 0.1 μm . ECD electron-capture detector was used for detection (Sartarelli *et al.*, 2012).

Preparation of fish samples and extraction

The extraction and cleanup processes were carried out using the original QuEChERS (rapid, easy, inexpensive, effective, robust, and safe) technique (Anastassiades *et al.*, 2003) and (Lehotay *et al.*, 2005): Before being used, fish samples were thawed overnight at 4 °C. Two grams of blank fish muscle samples were mixed and fortified with 20 μL of each of the following multistandard solutions: 75, 50, 25 and 5mg/kg (containing the three pesticides for 30 minutes). A 50 mL Falcon tube was filled with the combined meat samples. After adding 1.6 g of anhydrous MgSO_4 , 0.4 g of NaCl, and 4 mL of ACN (containing 1% acetic acid), the mixture was stirred for one minute in a vortex mixer. To separate the phases, the mixture was centrifuged for three minutes at 3000 rpm (liquid-liquid partition). The organic solvent ACN-corresponding supernatant was put into a tube with 150 mg of MgSO_4 , 70 mg of C_{18} and 70 mg of the adsorbent PSA. After a 30-second handshake, the tube was centrifuged for one minute at 4000 rpm. To eliminate the extra coloring components, the supernatant was filtered using a syringe filter (0.45 μm). The filtered supernatant was kept at 4 °C after being evaporated at 45 °C under a nitrogen stream. Under ideal circumstances, the solutions were moved to an auto sampler vial and injected into the GC-ECD.

The following equation was used to quantify the residues of the three studied pesticides:

$$\text{Concentration calculation } \left(\frac{\mu\text{g}}{\text{kg}} \right) = \frac{\text{Area of sample}}{\text{Area of standard}} * \text{concentration of standard} * \text{Dilution factor}$$

Table 1. Retention times (minutes) and area μvolt for three studied pesticides

Seq.	Type of pesticide	Retention time minute	Area μvolt
1	Hexachlorobenzene	3.032	1006356
2	Chlorpyrifos	4.43	913628
3	Cypermethrin	5.66	1042639

Method of validation

The globally recognized (SANTE, 2021) criteria-selectivity, recovery percentages, linearity, and sensitivity were followed in the validation process for this investigation. Five separate fish muscle sample extracts were injected into a GC-ECD detector to assess the method's selectivity. Each analyte's retention time window of interest was examined for the presence of interfering peaks above a signal-to-noise ratio of 3. Analytical standards (100 $\mu\text{g kg}^{-1}$) were injected into the GC-ECD apparatus individually in order to determine the target retention periods of the analytes. To enhance analytical validation and verify the range of retention time reproducibility, multi-standard solutions were introduced into the GC-ECD.

Recovery of fish muscle

By injecting multi-standard solutions containing three analytes at concentrations of 5, 25, 50 and 75 µg/kg for control matrices, recovery was assessed by comparing the obtained concentrations with the same concentrations of the pesticides prepared in the extraction solvent and analysing all spiked levels. To achieve the best technique reliability during screening, standard solutions were injected into blank matrices since each analyte in meat has a distinct maximum residue limit (MRL) (Nilsen, 2010).

Recovery studies for water samples

Distilled water samples were spiked with known amounts of pesticide and the spiked concentration sample and three blank samples were processed through the analytical method.

Limits of detection (LOD) and quantitation (LOQ)

Limits of detection and quantitation were determined based on a signal-to-noise ratio, and concentrations showing peak intensity of signal-to-noise ratio of 3 and 10 were designated as LOD and LOQ, respectively (SANTE, 2021).

Linearity

The linearity test in fish muscle was performed by injecting five matrix-matched standards for calibration purposes. Calibration curves and individual stock solutions of HCB, CPF, and CMN were made in Pyrex glass vials with acetonitrile concentrations of 100 µg kg⁻¹ and kept at -20 °C in dark amber bottles. Diluting the stock solution in acetonitrile yielded working standard solutions (WSS) with concentrations ranging from 5-75 µg kg⁻¹. Solutions were injected into extracted blank samples to achieve concentrations of 0.10 for each chemical produced, and linearity was assessed using calibration curve coefficients (r²) (SANTE, 2021).

Solid-phase extraction (SPE)

Solid-phase extraction (SPE) was used to remove pesticide residues from water samples (Kapsi *et al.*, 2020):

1. Before analysis, the water sample was vacuum-filtered via GF/B (1 µm glass fibre filters; Whatman, UK) to eliminate any suspended solid and prevent any potential interference.
2. Ten milliliters of acetone, ten milliliters of ethyl acetate, ten milliliters of methanol, and ten milliliters of deionised water were used to prepare the C18 discs.
3. Using a vacuum manifold that keeps the pressure differential between the disk's input and output constant, the 1000 mL of water samples were run through the SPE discs before the disc dried up at a flow rate of around 10 mL/min.
4. Rinse the discs with 2 x 5 mL of deionised water once the entire sample has percolated.
5. To get rid of any remaining water, the discs were vacuum-dried for ten minutes.
6. Nine milliliters of ethyl acetate/dichloromethane (85:15) were used to elute the analytes drop by drop at a rate of one milliliter per minute.
7. The excess anhydrous sodium sulphate was used to dry the final.
8. A mild stream of nitrogen caused the methanol extracts to evaporate until they were completely dry.
9. Before chromatographic analysis, reconstitute in 0.1 mL methanol and store at -20 °C.

Fat percentage

Fat in fish samples was estimated based on the AOAC method (2002) using a Soxhlet fat extraction apparatus in the presence of hexane as an organic solvent, by heating fish meat samples for 16 hours, and fat was extracted according to the following equation:

$$\text{Fat percentage} = \frac{\text{weight of flask before extraction} - \text{weight of flask after extraction}}{\text{weight of sample}} \times 100$$

Fish length

Fish specimen's total, fork and standard lengths were measured and documented using the SAP method. To be in line with the measure used by the majority of US fisheries scientists, total length should be applied (EPA, 2000). The subsequent techniques for measuring fish length using a measuring board were adopted from the United States Geological Survey (2002).

Calculating total length:

1. Position the fish so that its head is on the observer's left, its mouth was closed, and its body was on the right board.
2. Press the nose, or front end, up against the stop of the measurement board from mouth to caudal fin.
3. Determine the overall length by measuring the distance between the closed mouth and the caudal fin's extreme tip while its lobes were compressed dorsoventrally, or unsprayed.
4. To the closest millimeter, note the entire length.

Heat treatment

Heat treatment (grilling at 180 °C) was carried out using a Binder oven, made in Germany, for the muscles of studied fish for 30 minutes (Al-Joumaa, 2010).

Results and Discussions

Linearity range, regression equation

Table 2 shows five calibration curves (5-75 $\mu\text{g kg}^{-1}$) for the chemicals evaluated. The integrated peak area was plotted against the concentration. To ensure the linearity of calibration graphs, the correlation coefficient (R^2) for each chemical was computed. The calibration data of the pesticides investigated revealed high linearity for the response of the ECD detector, with R^2 values ranging from 0.9874 to 0.9996.

Table 2. Linearity range, regression equation coefficient

Pesticides	Linearity range ($\mu\text{g kg}^{-1}$)	Regression equation	R^2	LOD ($\mu\text{g kg}^{-1}$)	LOQ ($\mu\text{g kg}^{-1}$)
Hexachlorobenzene HCB	5-75	$Y=9877X+15800$	0.9993	14	42
Chlorpyrifos CPF	5-75	$Y=9563X+15062$	0.9874	6.0	18
Cypermethrin CMN	5-75	$Y=10339X+869.3$	0.9996	14	42

LODs, LOQs and Sensitivity

As shown in Table 2, the LOD of studied pesticides ranged from 6 to 14 $\mu\text{g kg}^{-1}$ in fish muscles, and the LOQ ranged from 18 to 42 $\mu\text{g kg}^{-1}$ in fish muscle samples (Table 2).

Recovery of fish muscle samples

The accuracy of the employed analytical method was determined via the calculation of average percentage recoveries for studied pesticides and the percent relative standard deviation (RSD %) of recoveries from fortified blank samples of fish (Table 3). The average percentages of recoveries and the RSD% of recoveries at 75 $\mu\text{g kg}^{-1}$ level of pesticide standards ranged from 99.86 ± 1.3 to $118.46 \pm 3.5\%$ and $97.90 \pm 3.2\%$ from spiked muscle samples for hexachlorobenzene, chlorpyrifos and cypermethrin, respectively. At 5 $\mu\text{g kg}^{-1}$ level of the reference material, the recovery percentages ranged from 104.6 ± 2.2 to $105.86 \pm 4.1\%$ and $99.6 \pm 3.2\%$ from spiked muscle samples for hexachlorobenzene, chlorpyrifos and cypermethrin, respectively (Table 3).

The findings demonstrated that the examined pesticide recovery rates were greater than 90% for every tissue sample. These findings are in line with the manual's (SANTE, 2021) permitted recovery rates for pesticide residues. The recovery rate should be between 70 and 120 percent, as numerous researchers have suggested (Stoytcheva, 2011). Recovery rates were in line with research by Fang *et al.*, (2024), which revealed that at a spiking level of 50 mg/kg, the tissue recovery rate of cypermethrin was 87.4%. At a dose of 30 $\mu\text{g kg}^{-1}$, the chlorpyrifos insecticide recovered at a rate of $81.32 \pm 2.876\%$ in normal carp tissues (Qayoom *et al.*, 2024). For fish (*Heteroclaris sp.*), the recovery rate of the pesticide hexachlorobenzene at a spiked level was 11 (ng g^{-1}) ($107.5\% \pm 3.8$) (Lourencetti and Ricci, 2020).

Table 3. Recovery of pesticides at different concentration spiked to blank compared with pure standard at 100 $\mu\text{g kg}^{-1}$ for fish tissue samples from Dukan Lake

Concentration (n $\mu\text{g kg}^{-1}$)	HCB		CPF		CMN	
Added	Found	Recovery (%)	Found	Recovery (%)	Found	Recovery (%)
75.0	74.899	99.86	88.1	118.46	73.43	97.90
50.0	49.92	99.84	52.99	105.9829	26.11	104.44
25.0	26.56	106.24	29.25	117.62	56.63	116.48
5.0	5.23	104.6	5.29	105.86	4.98	99.6

Recovery percentage for water samples

As can be seen in Table 4, from the 3 studied pesticides, there was an acceptable recovery percentage in the range from 70 to 120% for spiked levels 5, 25, 50 and 75 $\mu\text{g kg}^{-1}$, respectively.

Chlorpyrifos level from water in El-Behera was 2.25 ± 0.44 , in El- Qalyubia was 11.46 ± 1.4 and in El-Dakahlia was 7.10 ± 2.49 ppb in El-Gharbia was not detected (Nazeeh *et al.*, 2024). The level of chlorpyrifos residues in water samples from river Challawa was 35 to 0.78 $\mu\text{g kg}^{-1}$, while the concentrations of cypermethrin ranged from 0.01 to 0.09 mg L^{-1} (Akan *et al.*, 2014).

Table 4. Recovery of pesticides at different concentrations spiked to blank compared with pure standard at 100 µg L⁻¹ for water samples from Dukan Lake

Concentration (n µg/kg)	HCB		CPF		CMN	
Added	Found	Recovery %	Found	Recovery %	Found	Recovery%
75.0	77.96	103.94	76.98	102.64	74.23	98.97
50.0	50.57	101.14	51.73	103.46	57.13	114.26
25.0	26.56	106.24	29.25	117.62	25.56	102.24
5.0	4.7	94	5.29	105.86	4.85	97

Fish length

The length (total, standard, and fork) of 40 samples of common carp and barbus grypys was summarized in Table 5. Total lengths for common carp samples ranged from 30.5±0.40 cm to 33.83±1.30 cm, while total lengths for barbus grypys samples ranged from 36.5±0.23 to 40±2.16 cm.

Table 5. Mean of fish length (cm ± SE) samples were collected from Dukan Lake during different seasons

Seasons	Common carp			Barbus grypys		
	Total length (cm)	Standard length (cm)	Fork length (cm)	Total length (cm)	Standard length (cm)	Fork length (cm)
Autumn	31.5 ±0.47	28.16 ±0.59	26 ±0.47	40 ±2.16	37.16 ±2.29	34.66±1.9
Winter	30.5 ±0.40	27.33 ±0.36	24.83 ±0.36	39.16 ±0.49	36.66 ±0.72	34.5±0.62
Spring	30.33±1.20	27.5±1.04	25.33±0.93	36.5 ±0.23	32.66±0.27	31 ±0.0
Summer	33.83±1.30	30.63±1.35	28.16±1.17	38.5±1.80	35±1.89	32.6±2.04

Fish weight and fat percentage

Table 6 shows the weight of studied fish, weight of common carp from Dukan Lake ranged from 426±11.75 g to 523±7.55 g for different seasons, weight of barbus grypys ranged from 429±2.83g to 526±5.34 g for different seasons. The mean content of muscular fat ranged from 2.764±0.90 to 4.736 ±0.44% for common carp, while fat ranges for barbus grypys were between 2.354±1.01% to 3.701±1.14 %. The lowest mean content of fat in the muscles of common carp and barbus grypys was in winter while the highest value in autumn.

The findings of Yeganeh *et al.* (2012) (1.5%-5.1%), Čirković *et al.* (2012) (2.42% fat), Afkhami *et al.*, (2011) (3.53% fat), Čirković *et al.* (2011) (11.73% fat), and Marcu *et al.* (2010) (5.07% fat) were all consistent with the results of our investigation into the chemical composition of common carp. Regarding the fat content in the muscle of shabout fish in Dukan Lake that ranged from 2.354±1.01% to 3.372±0.74 % for different seasons, there was a significant difference among winter season with other seasons, fat percentage in winter recorded lower percentage. During sexual maturation, so the final stages of gonadal growth are dependent upon the mobilization and re-allocation of endogenous reserves. In the White Sea bream (*Diplodus sargus*), muscular lipid was highly mobilized during the spawning period, presumably in support of the reproductive effort (Pérez *et al.*, 2007).

Table 6. Mean of fish weight (g \pm SE) and fat % of muscle samples were collected from Dukan Lake during different seasons

Seasons	Common carp		Barbus grypus	
	Weight (g)	Fat %	Weight (g)	Fat %
Autumn	523 \pm 7.55	4.736 \pm 0.44	473 \pm 6.34	3.571 \pm 0.74
Winter	439 \pm 7.79	3.277 \pm 0.90	526 \pm 5.34	3.893 \pm 1.01
Spring	426 \pm 11.75	2.154 \pm 0.31	429 \pm 2.553	2.593 \pm 0.12
Summer	442 \pm 12.45	5.135 \pm 0.70	436 \pm 11.68	4.224 \pm 1.14

Pesticide residues from fish muscles

The residue concentrations of pesticides were detected in all seasons except hexachlorobenzene, which was not detected in spring from both fish species (Table 7). *Barbus grypus* exhibited the highest number of chlorpyrifos pesticide ($7.679 \pm 2.55 \mu\text{g kg}^{-1}$) in spring, however, the lowest value was detected in summer from common carp which was ($0.488 \pm 0.39 \mu\text{g kg}^{-1}$). Significant differences among seasons and types of were seen ($p \leq 0.05$).

The mean concentrations of chlorpyrifos in the muscle tissues of *S. melaanothern* were $0.0001 \pm 0.00002 \text{ mg kg}^{-1}$ in Etsii Lagoon and $0.0003 \pm 0.00003 \text{ mg kg}^{-1}$ in Fosu Lagoon, Ghana. The levels of pesticide residues found in the fish samples from Etsii and Fosu lagoons were all lower than those from their corresponding water samples (Essumang *et al.*, 2009). The buildup of chlorpyrifos at such high concentrations in the muscles of carp juveniles may be due to the chemical's structural composition (Qayoom *et al.*, 2024). Three chlorine atoms are present in the 2-ortho and 1-meta positions of the benzene ring. Consequently, the molecule is more like a compound of the organochlorine class than an organophosphate group. To put it another way, chlorpyrifos tends to accumulate in fish tissue at higher quantities than any other pesticide in its class and is more lipophilic than other organophosphates.

According to Zhao *et al.* (2011), who examined 30 fish tissue samples from a local fish market, one muscle sample tested positive for cypermethrin at $5.4 \mu\text{g kg}^{-1}$ and one liver sample tested positive at $7.2 \mu\text{g kg}^{-1}$ from crucian carp in China. Cypermethrin was discovered in seven of the 18 Canadian farmed salmon samples (39%), but not in any wild domestic salmon. Cypermethrin values in the positive samples varied between 0.3 and 6.5 ng g^{-1} . It was not found in any imported or domestically manufactured fish product (Rawn *et al.*, 2010). Mahmoob *et al.* (2015) studied the concentration of cypermethrin pesticide in the muscle tissues of different weight groups of *Catla catla* in dry and wet seasons. The dry season concentration ranged from 0.29 ± 0.06 to $3.45 \pm 0.10 \text{ ng g}^{-1}$, while the wet season's concentration ranged from undetectable to $2.85 \pm 0.36 \text{ ng/g}$.

Research has documented variations in HCB concentrations across quite narrow geographical areas. Because the East and West Coasts get air masses from various portions of North America, it has been demonstrated that the levels of HCB in Greenland polar cod are much greater on the East Coast than on the West Coast (Cleemann *et al.*, 2000). According to Muir *et al.* (1990), burbot livers in high-latitude locations in Canada have greater quantities of HCB, which may indicate that cold condensation is taking place. Fish and shellfish in the Belgian North Sea had lower levels of HCB the further they were from Antwerp, suggesting a higher amount of exposure upstream, according to Voorspoels *et al.* (2004). Some of the Organochlorine pesticides were higher in the wet season in the water, gills and muscle compared to dry season. This could be due to increased farming activities in the wet season, since farming is one of the primary occupations of the people that live around the water body (Abubakar *et al.*, 2024).

Table 7. Detected pesticide residues ($\mu\text{g kg}^{-1} \pm \text{SE}$) in common carp and barbus grypus muscles in Dukan Lake during different seasons

Seasons	Common carp			Barbus grypus		
	HCB	CPF	CMN	HCB	CPF	CMN
Autumn	1.170 \pm 0.85 ab	1.070 \pm 0.5 b	3.533 \pm 0.50 ab	2.087 \pm 0.96 ab	3.180 \pm 0.45 ab	4.033 \pm 0.19 ab
Winter	4.761 \pm 0.87 ab	2.871 \pm 1.09 ab	3.267 \pm 1.41 ab	5.815 \pm 1.82 ab	1.894 \pm 0.82 ab	1.397 \pm 0.34 ab
Spring	ND	4.415 \pm 1.23 ab	1.959 \pm 0.37 ab	ND	7.679 \pm 2.55 a	2.381 \pm 0.97 ab
Summer	0.488 \pm 0.39 b	0.806 \pm 0.38 b	0.662 \pm 0.54 b	2.657 \pm 1.07 ab	2.629 \pm 0.71 ab	1.356 \pm 0.55 ab

ND= Not detected

Pesticide residues from water

The results of the determination of pesticides in the surface water samples collected from Dukan Lake were summarized in Table 8. Pesticides were detected in all seasons, except hexachlorobenzene was not detected in summer. There were significant differences ($p \leq 0.05$) among seasons and types of pesticides, the highest residues were in autumn for chlorpyrifos 2.793 \pm 0.97 $\mu\text{g L}^{-1}$ while the lowest average was detected in spring, which was 0.277 \pm 0.13 $\mu\text{g L}^{-1}$.

Table 8. Detected pesticide residues ($\mu\text{g L}^{-1} \pm \text{SE}$) in water from Dukan Lake during different seasons

Seasons	Hexachlorobenzene ($\mu\text{g L}^{-1}$)	Chlorpyrifos ($\mu\text{g L}^{-1}$)	Cypermethrin ($\mu\text{g L}^{-1}$)
Autumn	0.790 \pm 0.21 abc	2.793 \pm 0.97 a	2.170 \pm 0.22 ab
Winter	2.291 \pm 0.46 ab	1.099 \pm 0.48 abc	0.478 \pm 0.23 bc
Spring	0.277 \pm 0.13 bc	1.025 \pm 0.42 abc	0.445 \pm 0.24 bc
Summer	ND	0.543 \pm 0.34 bc	1.222 \pm 0.99 abc

ND=Not detected.

Heat treatment

The effect of heat treatment on pesticide residues is presented in (Table 9). Grilled temperature reduced the pesticide residues by different percentages in both fish species and ranged from 26.01% to 43.39 % for three studied pesticides. Zabik *et al.* (1995) observed that heat treatment of Chinook salmon (*Oncorhynchus tshawytscha*) and common carp (*Cyprinus carpio*) reduced HCB, dieldrin, and total DDT residues by 30-41%. Hot smoking caused the greatest (over 50%) loss of HCB, total DDT, and chlordane complex in the muscle tissue of lake trout (*Salvelinus namaycush*) when exposed to various heat treatment procedures. In comparison, boiling or grilling were less successful, reducing pesticide concentration by 12-38 % on average (Zabik *et al.*, 1996).

Table 9. Effect of heat treatment (grilled at 180 °C for 30 minutes) on pesticide residues ($\mu\text{g kg}^{-1}$) in common carp and barbus grypus muscles in Dukan Lake during different seasons

Seasons	Types of fish	Type of pesticides	Raw muscle	Heat treated	Reduction rate (%)
Autumn	C. Carp	Hexachlorobenzene	3.533 \pm 0.50	2.493 \pm 0.11	29.66
		Chlorpyrifos	1.070 \pm 0.5	0.730 \pm 0.34	31.15
		Cypermethrin	1.170 \pm 0.85	0.809 \pm 0.59	30.79
	B. grypus	Hexachlorobenzene	4.033 \pm 0.19	2.83 \pm 1.02	33.42
		Chlorpyrifos	3.180 \pm 0.45	2.029 \pm 0.23	35.03
		Cypermethrin	2.087 \pm 0.96	1.516 \pm 0.71	27.33
Winter	C. Carp	Hexachlorobenzene	3.267 \pm 1.41	2.41 \pm 1.01	26.34
		Chlorpyrifos	2.871 \pm 1.09	1.950 \pm 0.78	31.93
		Cypermethrin	4.761 \pm 0.87	3.055 \pm 0.48	35.82
	B. grypus	Hexachlorobenzene	1.397 \pm 0.34	1.033 \pm 0.85	26.01
		Chlorpyrifos	1.894 \pm 0.82	1.322 \pm 0.57	32.82
		Cypermethrin	5.815 \pm 1.82	4.117 \pm 1.02	29.18
Spring	C. Carp	Hexachlorobenzene	1.959 \pm 0.37	1.56 \pm 0.29	20.33
		Chlorpyrifos	4.415 \pm 1.23	1.666 \pm 0.37	43.39
		Cypermethrin	ND	ND	
	B. grypus	Hexachlorobenzene	2.381 \pm 0.97	1.555 \pm 0.64	34.58
		Chlorpyrifos	7.679 \pm 2.55	5.343 \pm 1.45	30.41
		Cypermethrin	ND	ND	
Summer	C. Carp	Hexachlorobenzene	0.662 \pm 0.54	0.431 \pm 0.35	34.87
		Chlorpyrifos	0.806 \pm 0.38	0.533 \pm 0.26	31.94
		Cypermethrin	0.488 \pm 0.39	0.348 \pm 0.28	28.50
	B. grypus	Hexachlorobenzene	1.356 \pm 0.55	0.792 \pm 0.32	43.54
		Chlorpyrifos	2.629 \pm 0.71	1.850 \pm 0.22	30.14
		Cypermethrin	2.657 \pm 1.07	1.546 \pm 0.35	41.79

ND= Not detected.

Comparing pesticide residues in fish tissues with MRLs

Hexachlorobenzene, chlorpyrifos and cypermethrin were compared with pesticide MRL data for seafood from Japan and the European Union (EU), (Table 10). Japan and European Union have established standards for all three pesticides. Japan has set MRL for hexachlorobenzene, chlorpyrifos and cypermethrin (100, 300, 50 $\mu\text{g kg}^{-1}$), respectively. There were no samples that exceeded the Japan MRL. European Union has established a MRL for pesticides in fish; only one sample exceeded the EU MRL which was in winter and from b. grypus (11.964 $\mu\text{g kg}^{-1}$).

Table 10. Comparison of pesticide residues from fish muscle with maximum residue limits (MRL) standard of pesticides in Japan and EU ($\mu\text{g kg}^{-1}$)

Seasons	Types of fish	Type of pesticides	Range ($\mu\text{g kg}^{-1}$)	MRL EU	MRL Japan
Autumn	C. Carp	Hexachlorobenzene	0-6.19	10	100
		Chlorpyrifos	0-2.13	10	300
		Cypermethrin	0-3.25	10	50
	B. grypus	Hexachlorobenzene	3.55-4.28	10	100
		Chlorpyrifos	2.08-3.89	10	300
		Cypermethrin	0-4.07	10	50
		Hexachlorobenzene	0-5.23	10	100

Winter	C. Carp	Chlorpyrifos	0.287-4.788	10	300
		Cypermethrin	3.458-6.874	10	50
	B. grypus	Hexachlorobenzene	0-4.19	10	100
		Chlorpyrifos	0-3.439	10	300
		Cypermethrin	0-11.964	10	50
Spring	C. Carp	Hexachlorobenzene	0-5.876	10	100
		Chlorpyrifos	0-8.830	10	300
		Cypermethrin	ND	10	50
	B. grypus	Hexachlorobenzene	0-3.633	10	100
		Chlorpyrifos	0-19.296	10	300
		Cypermethrin	ND	10	50
Summer	C. Carp	Hexachlorobenzene	0-1.986	10	100
		Chlorpyrifos	0-1.627	10	300
		Cypermethrin	0-1.463	10	50
	B. grypus	Hexachlorobenzene	0-2.182	10	100
		Chlorpyrifos	0-6.763	10	300
		Cypermethrin	0-7.971	10	50

ND= Not detected, EU=European Union.

Pesticide residues from water with MRLs

The residue levels of the detected pesticides in this monitoring were found to be considerably lower than foreign standards and compared with Japan and USA MRLs (Table 11). Japan only established MRL for chlorpyrifos for mineral/packaged drinking water whereas USA established MRL for hexachlorobenzene ($30 \mu\text{g L}^{-1}$) and chlorpyrifos ($50 \mu\text{g L}^{-1}$).

Table 11. Comparison of pesticide residues from water with maximum residue limits (MRL) standard of pesticides in Japan and USA ($\mu\text{g L}^{-1}$)

Seasons	Type of pesticides	Range	MRL USA	MRL Japan
Autumn	Hexachlorobenzene	1.73-2.69	30	-
	Chlorpyrifos	1.35-5.43	50	30
	Cypermethrin	0.32-1.21	-	-
Winter	Hexachlorobenzene	0-0.987	30	-
	Chlorpyrifos	0.18-2.205	50	30
	Cypermethrin	1.41-3.298	-	-
Spring	Hexachlorobenzene	0-1.017	30	-
	Chlorpyrifos	0-1.659	50	30
	Cypermethrin	0-0.581	-	-
Summer	Hexachlorobenzene	0-3.664	30	-
	Chlorpyrifos	0-1.627	50	30
	Cypermethrin	ND	-	-

ND= Not detected.

Correlation matrix (Pearson) among parameters

Fish muscles from same sampling zone showed no statistical differences ($p \leq 0.05$) in each instance except the correlation among cypermethrin and chlorpyrifos. Correlation analysis revealed a positive correlation among hexachlorobenzene residue concentration, fish weight, fish length and fat percentage (Table 12).

According to Akor *et al.* (2021), lindane and p,p'-DDT were found to be positively correlated with fat in various organs of the sampled fish species, which supports the findings of Dang *et al.* (2016) who discovered that the bioaccumulation of certain types of contaminants in animal tissues is generally proportional to their lipid content. There was a high positive association between lindane and fat in catfish liver and muscle tissue, indicating a direct proportional link; hence, as fat levels grew, so did lindane accumulation in the liver and muscle tissue. This shows that excessive fat content is a significant contributor to lindane buildup in liver and muscle tissue, especially given the high overall prevalence rate of 63.8% in all samples analysed. This is consistent with the literature as stated by Helberg *et al.* (2005).

Table 12. Correlation matrix (Pearson) among parameters

Variables	Fish weight	Fish length	Fat percentage	HCB	CMN	CPF
Fish weight	1					
Fish length	0.391	1				
Fat percentage	0.341	0.129	1			
HCB	0.193	0.151	0.127	1		
CMN	-0.050	-0.062	-0.001	0.128	1	
CPF	-0.160	0.159	-0.205	-0.032	0.422	1

Values in bold are different from 0 with a significance level $\alpha=0.05$.

Conclusions

This study showed that pesticide residues were found in both fish tissues and water from Dukan Lake. However, the degree of pollution did not pose any significant health hazards to all of the investigated fish species. Nonetheless, there should be a public awareness campaign about the existence of these toxins in the Lake's fish. As a result, farmers in the study region should apply pesticides far enough away from the lake to safeguard aquatic life and the health of humans who consume polluted fish.

Conflict of interest

The authors confirm that there are no conflicts of interest to declare regarding the publication of this paper.

Acknowledgments

We express our gratitude to everyone who helped us to complete this research.

References

- Abubakar, I., Dasuki, A., Babatunde, T. A., Ghali-Mohammed, I., Abdurrasheed, N., and Dauda, A. B. (2024). Seasonal variation of organochlorine pesticides residue in water and silver catfish (*Bagrus Bajad fabricius*, 1775) from Ajiwa Reservoir, Katsina State. *Journal of Research in Forestry, Wildlife and Environment*, 16(2), 9-16. <https://www.ajol.info/index.php/jrfwe/article/view/275304>
- Afkhami, M., Mokhlesi, A., Bastami, K. D., Khoshnood, R., Eshaghi, N., and Ehsanpour, M. (2011). Survey of some chemical compositions and fatty acids in cultured common carp (*Cyprinus carpio*) and grass carp (*Ctenopharyngodon idella*), Noshahr, Iran. *World Journal of Fish and Marine Sciences*, 3(6), 533-538. <https://doi.org/10.13140/2.1.4317.2640>

- Akan, J. C., Sodipo, O. A., Mohammed, Z., and Abdulrahman, F. I. (2014). Determination of organochlorine, organophosphorus and pyrethroid pesticide residues in water and sediment samples by high performance liquid chromatography (HPLC) with UV/visible detector. *Journal of Analytical & Bioanalytical Techniques*, 5(6), 1-5. <https://doi.org/10.4172/2155-9872.1000226>
- Akor, A. O., Faruruwa, M. D., and Mohammed, Y. (2021). Assessment of Organochlorine Pesticide Residue Levels and Fat Content in Liver, Gill and Muscle Tissues of Catfish (*Clarias Spp*) and Tilapia (*Oreochromis Spp*) Obtained from River Kaduna and Fish Farms in Kaduna Metropolis, Nigeria. *IOSR Journal of Environmental Science, Toxicology and Food Technology*, 15(12), 26-36. <https://www.iosrjournals.org/iosr-jestft/papers/Vol15-Issue12/Ser-1/D1512012636.pdf>
- Al-Joumaa, K. (2010). Effect of heat treatment on the nutritive value and residues of some synthetic pesticides in fresh Bolti fish, *Arab Universities Journal of Agricultural Sciences*, 18(2), 329-335. <https://doi.org/10.21608/ajs.2010.14900>
- Anastassiades, M., Lehotay, S. J., Štajnbaher, D., and Schenck, F. J. (2003). Fast and easy multiresidue method employing acetonitrile extraction/partitioning and “dispersive solid-phase extraction” for the determination of pesticide residues in produce. *Journal of AOAC international*, 86(2), 412-431. <https://doi.org/10.1093/jaoac/86.2.412>
- AOAC. (2002). *Association of Official Analytical Chemists, official method of analysis*, 17th ed. A.O.A.C. International, Washington. D.C, USA.
- Rajendran, R. B., Imagawa, T., Tao, H., and Ramesh, R. (2005). Distribution of PCBs, HCHs and DDTs, and their ecotoxicological implications in Bay of Bengal, India. *Environment International*, 31(4), 503-512. <https://doi.org/10.1016/j.envint.2004.10.009>
- Botaro, D., Torres, J. P. M., Malm, O., Rebelo, M. F., Henkelmann, B., and Schramm, K. W. (2011). Organochlorine pesticides residues in feed and muscle of farmed Nile tilapia from Brazilian fish farms. *Food and Chemical Toxicology*, 49(9), 2125-2130. <https://doi.org/10.1016/j.fct.2011.05.027>
- Burkepile, D. E., Moore, M. T., and Holland, M. M. (2000). Susceptibility of five nontarget organisms to aqueous diazinon exposure. *Bulletin of Environmental Contamination and Toxicology*, 64(1). <https://doi.org/10.1007/s001289910018>
- Čirković, M., Ljubojević, D., Tupan, B., Bogut, I., Đordjević, V., Novakov, N., and Matekalo – Sverak, V. (2012). Usporedni prikaz kvalitete mesa nekih vrsta riba iz porodice šaranki u Republici Srbiji. *Croatian Journal of Fisheries*, 70 (1), 79-88.
- Čirković, M., Trbović, D., and Ljubojević, D. (2011). Meat quality of fish farmed in polyculture in carp ponds in Republic of Serbia. *Meat Technology*, 52, 106-121. http://www.journalmeattechnology.com/index.php/meat_technology/article/view/264
- Cleemann, M., Riget, F., Paulsen, G. B., Klungsøyr, J., and Dietz, R. (2000). Organochlorines in Greenland marine fish, mussels and sediments. *Science of the Total Environment*, 245(1-3), 87-102. [https://doi.org/10.1016/S0048-9697\(99\)00435-0](https://doi.org/10.1016/S0048-9697(99)00435-0)
- Dang, V. D., Kroll, K. J., Supowit, S. D., Halden, R. U., and Denslow, N. D. (2016). Tissue distribution of organochlorine pesticides in largemouth bass (*Micropterus salmoides*) from laboratory exposure and a contaminated lake. *Environmental Pollution*, 216, 877-883. <https://doi.org/10.1016/j.envpol.2016.06.061>
- EPA, U. (2000). Guidance for assessing chemical contaminant data for use in fish advisories. Risk assessment and fish consumption limits, 2. <https://www.epa.gov/sites/default/files/2015-06/documents/volume2.pdf>

- Essumang, D. K., Togoh, G. K., and Chokky, L. (2009). Pesticide residues in the water and fish (lagoon tilapia) samples from lagoons in Ghana. *Bulletin of the Chemical Society of Ethiopia*, 23(1), 19-27. <https://doi.org/10.4314/bcse.v23i1.21294>
- Fang, Changling, Xiaoyi Lou, Xuan Zhang, Siman Li, Yunyu Tang, Yongfu Shi, and Dongmei Huang. (2024). Simultaneous Determination of Seven Pyrethroid Pesticide Residues in Aquatic Products by Gas Chromatography" *Fishes*, 9(3), 1-13. <https://doi.org/10.3390/fishes9030079>
- Hassan, S. M., Albassam, N. H., Madlul, N. S., and Mahmood, A. S. (2024). Influence of different salinity concentrations on water quality and subsequent effects on hatchability rate of Artemia salina eggs. *AIP Conference Proceedings*, 3079, (1), 020007. <https://doi.org/10.1063/5.0202200>
- Hassan, S.M., Rashid, M.S., Albassam, N.H., and Sulaiman, M. A. (2025). Using the Diverse Vegetables as a Filtration plants in Aquaculture Intensive System. *Tikrit Journal for Agricultural Sciences*, 25 (1), 1–16. <https://doi.org/10.25130/tjas.25.1.1>
- Helberg, M., Bustnes, J. O., Erikstad, K.E., Kristiansen, K. O. and Skaare, J. U. (2005). Relationships between Reproductive Performance and Organochlorine Contaminants in Great Black-backed Gulls (*Larus marinus*). *Environmental Pollution*. 134, 475-483. <https://doi.org/10.1016/j.envpol.2004.09.006>
- Jiries, A. G., Al Nasir, F. M., and Beese, F. (2002). Pesticide and heavy metals residue in wastewater, soil and plants in wastewater disposal site near Al-Lajoun Valley, Karak/Jordan. *Water, Air, and Soil Pollution*, 133, 97-107. <https://doi.org/10.1023/A:1012923832506>
- Kapsi, M., Tsoutsis, C., and Albanis, T. (2020). Simple analytical methodology based on solid phase extraction for monitoring pesticide residues in natural waters. *MethodsX*, 7, 1-8. <https://doi.org/10.1016/j.mex.2020.101011>
- Kumari, A., Srivastava, A., and Jha, M. M. (2014). Carbaryl Induced alteration in histology and certain biochemical parameter in liver of Clarias batrachus. *Global Journal of Bioscience and Biotechnology*, 3(3), 259-263.
- Lehotay, S. J., Kok, A. D., Hiemstra, M., and Bodegraven, P. V. (2005). Validation of a fast and easy method for the determination of residues from 229 pesticides in fruits and vegetables using gas and liquid chromatography and mass spectrometric detection. *Journal of AOAC International*, 88(2), 595-614. <https://doi.org/10.1093/jaoac/88.2.595>
- Lourencetti, C., and Ricci, M. (2020). Determination of organochlorine priority substances in fish tissue: Optimisation of the clean-up step balancing removal of lipids with analytes' recovery. *Journal of Chromatography A*, 1619, 1-10. <https://doi.org/10.1016/j.chroma.2020.460944>
- Marcu, A., Ileana, N., Maria, N., Adrian, M., and Bartolomeu, K. (2010). Studies regarding the meat quality of the species Cyprinus carpio. *Lucrari Stiintifice Medicina Veterinara*, XLIII 2, 1-6. <https://www.cabidigitallibrary.org/doi/pdf/10.5555/20103231304>
- Maule, A. G., Gannam, A. L., and Davis, J. W. (2007). Chemical contaminants in fish feeds used in federal salmonid hatcheries in the USA. *Chemosphere*, 67(7), 1308-1315. <https://doi.org/10.1016/j.chemosphere.2006.11.029>
- Monirith, I., Nakata, H., Watanabe, M., Takahashi, S., Tanabe, S., and Tana, T. S. (2000). Organochlorine contamination in fish and mussels from Cambodia and other Asian countries. *Water Science and Technology*, 42(7-8), 241-252. <https://doi.org/10.2166/wst.2000.0575>
- Muir, D. C., Ford, C. A., Grift, N. P., Metner, D. A., and Lockhart, W. L. (1990). Geographic variation of chlorinated hydrocarbons in burbot (Lota lota) from remote lakes and rivers in

Canada. *Archives of Environmental Contamination and Toxicology*, 19, 530-542.
<https://doi.org/10.1007/BF01059072>

- Nazeeh, N., Radwan, E. H., Mosalam Hosin, E., and Khalifa, H. (2024). A comparative study of the effect of pesticides used in agriculture on catfish in some Nile Delta governorates. *Biological and Biomedical Journal*, 2(2), 168-180.
<https://www.ajol.info/index.php/bbj/article/view/290476>
- Nilsen, C. H. (2010). Chromatography of metabolites in plasma and urine following oral administration of anthocyanin rich capsules. *Master Thesis in Science and Technology. University of Stavanger, Norway*. <https://uis.brage.unit.no/uis-xmlui/handle/11250/182430>
- Ntow, W. J. (2005). Pesticide residues in Volta lakeLake, Ghana. *Lakes and Reservoirs: Research and Management*, 10(4), 243-248. <https://doi.org/10.1111/j.1440-1770.2005.00278.x>
- Pérez M.J., Rodriguez C., Cejas J.R., Martín M.V., Jerez S., and Lorenzo A. (2007). Lipid and fatty acid content in wild White sea bream (*Diplodus sargus*) broodstock in different stages of the reproductive cycle. *Comparative Biochemistry and Physiology Part B: Biochemistry and Molecular Biology*, 146(2), 187–196. <https://doi.org/10.1016/j.cbpb.2006.10.097>
- Qayoom, I., Balkhi, M., Mukhtar, M., Abubakr, A., Siddiqui, U., Khan, S., and Mastinu, A. (2024). Assessing organophosphate insecticide retention in muscle tissues of juvenile common carp fish under acute toxicity tests. *Toxicology Reports*, 12, 253-259.
<https://doi.org/10.1016/j.toxrep.2024.02.002>
- Rawn, D. F., Judge, J., and Roscoe, V. (2010). Application of the QuEChERS method for the analysis of pyrethrins and pyrethroids in fish tissues. *Analytical and Bioanalytical Chemistry*, 397, 2525-2531. <https://doi.org/10.1007/s00216-010-3786-5>
- SANTE, (2021). *Guidance document on analytical quality control and validation procedures for pesticide residues analysis in food and feed*. Supersedes Document No. SANTE/2019/12682.
- Sartarelli, N. C., de Macedo, A. N., de Sousa, J. P., Nogueira, A. R. D. A., and Brondi, S. H. G. (2012). Determination of chlorfenvinphos, fipronil, and cypermethrin residues in meat and bovine fat using QuEChERS method and gas chromatography-mass spectrometry. *Journal of Liquid Chromatography and Related Technologies*, 35(13), 1895-1908.
<https://doi.org/10.1080/10826076.2011.627609>
- Srivastava, A. K., Mishra, D., Shrivastava, S., Srivastav, S. K., and Srivastav, A. K. (2010). Acute toxicity and behavioural responses of Heteropneustes fossilis to an organophosphate insecticide, dimethoate. *International Journal of Pharma and Bio Sciences*, 1(4), 359-363.
- Storelli, M. M. (2008). Potential human health risks from metals (Hg, Cd, and Pb) and polychlorinated biphenyls (PCBs) via seafood consumption: estimation of target hazard quotients (THQs) and toxic equivalents (TEQs). *Food and Chemical Toxicology*, 46(8), 2782-2788.
<https://doi.org/10.1016/j.fct.2008.05.011>
- Stoytcheva, M. (Ed.). (2011). *Pesticides in the modern world: Risks and Benefits*. BoD–Books on Demand. <https://doi.org/10.5772/949>
- United States Geological Survey. (2002). Revised Protocols for Sampling Algal, Invertebrate, and Fish Communities as Part of the National Water-Quality Assessment Program. Reston, VA: U.S. Geological Survey. <https://doi.org/10.3133/ofr2002150>
- VanCuren, R. A. (2003). Asian aerosols in North America: Extracting the chemical composition and mass concentration of the Asian continental aerosol plume from long-term aerosol records in the western United States. *Journal of Geophysical Research: Atmospheres*, 108(20), 1-20.
<https://doi.org/10.1029/2003JD003459>

- Voorspoels, S., Covaci, A., Maervoet, J., De Meester, I., and Schepens, P. (2004). Levels and profiles of PCBs and OCPs in marine benthic species from the Belgian North Sea and the Western Scheldt Estuary. *Marine pollution bulletin*, 49(5-6), 393-404.
<https://doi.org/10.1016/j.marpolbul.2004.02.024>
- Yahia, D., and Elsharkawy, E. E. (2014). Multi pesticide and PCB residues in Nile tilapia and catfish in Assiut city, Egypt. *Science of the Total Environment*, 466, 306-314.
<https://doi.org/10.1016/j.scitotenv.2013.07.002>
- Yeganeh, S., Shabanpour, B., and Shabani, A. (2012). Comparison of farmed and wild common carp (*Cyprinus carpio*): Seasonal variations in chemical composition and fatty acid profile. *Czech Journal of Food Sciences*, 30(6), 503. <https://doi.org/10.17221/455/2011-CJFS>
- Zabik M.E., Zabik M.J., Booren A. M., Nettles M., Song J.-H., Welch R. and Humphrey H., (1995). Pesticides and total polychlorinated biphenyls in Chinook Salmon and Carp harvested from the great Lakes: Effects of skin-on and skin-off processing and selected cooking methods. *Journal of Agricultural and Food Chemistry*, 43, 993–1001.
<https://pubs.acs.org/doi/pdf/10.1021/jf00052a029>
- Zabik M.E., Booren A., Zabik M.J., Welch R. and Humphrey H. (1996). Pesticide residues, PCBs and PAHs in baked, charbroiled, saltboiled and smoked Great Lakes lake trout. *Food Chemistry*, 55(3), 231-237. [https://doi.org/10.1016/0308-8146\(95\)00115-8](https://doi.org/10.1016/0308-8146(95)00115-8)
- Zhao, D., Liu, X., Shi, W., and Liu, R. (2011). Determination of cypermethrin residues in crucian carp tissues by MSPD/GC-ECD. *Chromatographia*, 73, 1021-1025.
<https://doi.org/10.1007/s10337-011-1921-x>
- Zidan, Z. H., Elewa, I. S., Hussein, M. I., Mohamed, K. A., and Al-Naser, Z. A. (1997). Persistence and distribution of certain fungicides in soil and tomato plants. *Annals of Agricultural Science*, 42(2), 675-686.